



# Synthesis and photocatalytic activity of Mn-doped TiO<sub>2</sub> nanostructured powders under UV and visible light

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## ABSTRACT

A straight forward, simple and inexpensive process has been developed by sol–gel method for the synthesis of manganese (Mn) doped and undoped TiO<sub>2</sub> photocatalysts. X-ray powder diffractometry (XRD), scanning electron microscopy (SEM) and transmission electron microscopy (TEM) revealed the presence of structural nanoparticles with an average nanocrystalline size of about 20 nm. FT-IR/vis adsorption has shown enhanced sub-band-gap absorption as a function of Mn concentration. The photocatalytic activity of these materials was evaluated by the degradation of a basic Methylene Blue (MB) as organic contaminant. Mn-doped TiO<sub>2</sub> powder with molar ratio 0.1:100 was mixed with calcareous filler (5% and 10% respectively) and the photocatalytic activity was evaluated by the degradation of inorganic pollutants such as NO<sub>x</sub> under UV and visible light.

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## 1. Introduction

Titanium dioxide (TiO<sub>2</sub>) has been widely studied particularly since the discovery of its photocatalytic properties by Fujishima and Honda in 1972 [1]. Some of the current promising applications for TiO<sub>2</sub> nanomaterials include its content in paints [2], and toothpastes [3], while it is used also for UV protection, in photocatalysis [4,5], in photovoltaics, in gas sensing, in electrochromics as well as photochromics [6]. As a catalyst, TiO<sub>2</sub> nanomaterials are very stable, nontoxic, cheap and have high photocatalytic reactivity in the elimination of pollutants in air and water [7]. A very important application of TiO<sub>2</sub> photocatalysts is in the purification of indoor and outdoor air [8]. Malodorous substances such as ammonia, hydrogen sulfide, formaldehyde, acetaldehyde, toluene, etc., constitute potential risks to human health and comfort. Fortunately, these dangerous volatile organic compounds (VOCs) oxidize in the presence of TiO<sub>2</sub> nanomaterials into harmless compounds such as CO<sub>2</sub> and H<sub>2</sub>O under irradiation with UV and visible light [9–22]. The catalytic mechanism is based on the excitation of electrons from the valence band (VB) to the conduction band (CB) of the TiO<sub>2</sub>, leaving holes in the Valence Band. These electrons (e<sup>−</sup>) and

holes (h<sup>+</sup>) can then initiate redox reactions with molecular species adsorbed on the surface of the catalyst.

Titanium dioxide can have three different crystalline phases – *rutile* (R), *anatase* (A) and *brookite* (B). There are a large number of publications reporting on the effect of phase composition and particle size on the efficiency and the photocatalytic degradation of organic pollutants. Some demonstrate that *anatase* has a higher photocatalytic activity than *rutile*, while in others the best performance is related to *rutile* [23–27]. Many experimental data indicate that mixed phases, with small amount of *rutile*, can facilitate the separation of photogenerated charge carriers and this enhances the photocatalytic activity [28].

Anatase TiO<sub>2</sub> has an energy band gap of 3.2 eV and can be activated by UV radiation with a wavelength up to 387 nm thus limiting its application as a photocatalyst under visible light illumination. However, doped with transition metal ions, the TiO<sub>2</sub> matrix has been frequently reported to suppressing charge carrier recombination and facilitates the onset shift in the band gap absorption to the visible region, thus enhancing its photocatalytic activity [29]. Choi et al. performed a systematic study of TiO<sub>2</sub> nanoparticles doped with 21 metal ions by the sol–gel method and found that the presence of metal ion dopants significantly influenced the photoreactivity, charge carrier recombination rates, and interfacial electron-transfer rates [30]. Paola et al. have investigated the doping of Co, Cr, Cu, Fe, Mo, V and W into TiO<sub>2</sub> matrix prepared by wet impregnation method [31]. Nagaveni et al. reported an inhibitory

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effect for the dopants W, V, Ce, Zr, Fe and Cu prepared by the solution combustion method [32] while few researchers have claimed the enhanced activity for these dopants compared to undoped  $\text{TiO}_2$  in photocatalysis [33]. Kumar and co-workers have studied the photocatalytic activity of aniline blue under UV/solar light for the dopants  $\text{Mn}^{2+}$ ,  $\text{Ni}^{2+}$ , and  $\text{Zn}^{2+}$  into the  $\text{TiO}_2$  matrix [34].

In this work, we report on the synthesis and characterization of  $\text{TiO}_2$  nanomaterials doped with manganese (Mn) in different concentrations capable to absorb and activate under visible light irradiation. The Mn-doped  $\text{TiO}_2$  ( $\text{Mn}:\text{TiO}_2$ ) powders were prepared by a modified sol–gel method. The photocatalyst was synthesized [patent GR20090100724] following a few simple synthetic steps listed below with two different precursors namely titanium (IV) tetra-isopropoxide ( $\text{Ti}(\text{OCH}(\text{CH}_3)_2)_4$ ) and titanium (IV) oxysulfate ( $\text{TiOSO}_4 \cdot x\text{H}_2\text{O}$ ). Photocatalytic material may be obtained with the same physical properties, the same structure and almost the same photocatalytic activity regardless of the above precursors involved. However, the titanium (IV) tetra-isopropoxide is expensive and the synthesis requires organic solvents which increase the price of the final product; also the synthesis is harder to scale up, in contrast with the titanium (IV) oxysulfate ( $\text{TiOSO}_4 \cdot x\text{H}_2\text{O}$ ) precursor applied here. To improve the photocatalytic activities we doped  $\text{TiO}_2$  with manganese and carried out a study of its photocatalytic activity using UV and visible light as source of radiation with Methylene Blue (MB) as the test contaminant. These photocatalytic materials may be incorporated either into concrete surfaces or applied as coatings (i.e. in walls roofs and floors) for indoor depollution applications. They may also be applied in out-door applications (i.e. sidewalks, landscaping of buildings, parking lots, etc.), contributing to improved air quality and urban environments through the reduction of pollutant agents.

In the present work, we studied the photocatalytic activity of  $\text{Mn}:\text{TiO}_2$  powders through the degradation of a basic MB organic contaminant. Further on we mixed the above photocatalytic powders with calcareous filler base coatings.  $\text{Mn}:\text{TiO}_2$  powders alone as well as special coatings/overlays of calcareous base with 5% and 10% powder were prepared and studies were carried out on their photocatalytic activity using UV and visible light as source of radiation towards  $\text{NO}_x$ , as inorganic pollutant.

## 2. Experimental

### 2.1. Preparation of undoped and Mn doped $\text{TiO}_2$

Titanium (IV) oxysulfate hydrate ( $\text{TiOSO}_4 \cdot x\text{H}_2\text{O}$ ), manganese (II) acetate tetrahydrate,  $\geq 99\%$  and titanium (IV) tetraisopropoxide [ $\text{Ti}(\text{OCH}(\text{CH}_3)_2)_4$ ] purchased from Aldrich were applied. The Mn doped  $\text{TiO}_2$  was prepared by a modified sol–gel method. The photocatalyst was obtained by precipitating titanium dioxide on a sol of manganese dioxide, as shown in the flowchart of the synthesis employing  $\text{TiOSO}_4$  of Fig. 1. The hydrated manganese dioxide sol was obtained by mixing the required volumes of manganese acetate, i.e.  $\text{Mn}(\text{CH}_3\text{COO})_2$ , 0.1 M and potassium permanganate  $\text{KMnO}_4$ , 0.1 M solutions and stirring the mixture for 24 h at room temperature. The obtained sol was mixed with a solution of  $\text{TiOSO}_4$ . The concentration of  $\text{TiOSO}_4$  in the final solution was 0.1 M. The colloidal solution was stirred (Fig. 1/step1) at room temperature for 48 h, in order to obtain the adsorption equilibrium. During this phase an exchange of Mn with Ti occurs and the final sol was a mixture of both dioxides. After this step, the remaining  $\text{Ti}^{4+}$  ions were forced to precipitate by adding  $\text{NH}_3$  solution so that the final pH was 7 (step 2). Gel so formed was stirred continuously at room temperature for aging (48 h). After aging the sol was separated using centrifuging or alternatively it was filtered under vacuum, to obtain a powder (steps 3–4). The powder was then washed (step 5) with

distilled water until free of sulfate and ammonium ions. The powder was free of sulfate and ammonium ions when the test of sulfate and ammonium was negative. If the test was positive, the procedure (step 5) was repeated. The powders, after drying at  $100^\circ\text{C}$ , were calcined for 3 h at  $700^\circ\text{C}$ . Mn was loaded in  $\text{TiO}_2$  with molar ratio in different concentrations between 0 and 1 wt%.

Furthermore, samples of calcareous filler containing 0%, 5% or 10% of the 0.1%  $\text{Mn}:\text{TiO}_2$  photocatalyst were deposited on  $0.25\text{ m} \times 0.25\text{ m}$  glass panels.

### 2.2. Photocatalytic experiments

Photocatalytic degradation of MB ( $\text{C}_{16}\text{H}_{18}\text{ClN}_3\text{S}$ ) in aqueous solution ( $10\text{ mg L}^{-1}$ ) was carried out under UV (310–400 nm wavelength) using a mercury Lamp (Spectroline ultraviolet quartz pencil lamp/ $l = 4.5\text{ mW cm}^{-2}$ ) and visible-light (Osram Dulux, 11 W) irradiation. About 100 mg of photocatalyst was added into a 100 ml of aqueous MB (10 ppm) solution. Prior to irradiation, the suspensions were sonicated in the dark for 30 min to ensure establishment of an adsorption–desorption equilibrium among the photocatalyst, MB and atmospheric oxygen. After recovering the catalyst by centrifuging, subsequent variations in the MB concentration were analyzed by an Ultraviolet–visible spectrophotometer (VARIAN Cary 50) and the adsorption peak at 664 nm ( $\lambda_{\text{max}}$  for MB) was recorded.

The photocatalytic activity of calcareous paint (containing 5% and 10% of the 0.1 wt% of  $\text{Mn}:\text{TiO}_2$ ) was examined employing a photocatalytic system which was comprised of a gas delivery unit, a reactor chamber and a measurement unit (Fig. 2). A compressed gas cylinder of 10 ppmv NO ( $\pm 2\%$ ) balanced in  $\text{N}_2$  was used for the creation of  $\text{NO}_x$  polluted atmosphere in the photocatalytic reactor (Linde Hellas Ltd.). Prior to its introduction to the photocatalytic chamber, NO was mixed with the appropriate quantity of synthetic air (20.5% v/v  $\text{O}_2$  and 79.5% v/v  $\text{N}_2$ ) adjusting the initial  $\text{NO}_x$  concentration to  $C_{\text{NO}} = 250\text{ }\mu\text{g m}^{-3}$ .

Photocatalytic experiments were carried out in a cubic pyrex glass reaction chamber of  $0.125\text{ m}^3$  volume capacity ( $50\text{ cm} \times 50\text{ cm} \times 50\text{ cm}$ ) which was placed inside a (light) sealed irradiation box. The photocatalytic films were placed at the bottom surface of the cell. The cell's loading factor  $l_f$  (photocatalyst surface/cell volume ratio) was estimated to be about  $0.442\text{ m}^2\text{ m}^{-3}$ .

Illumination was provided by ten 15 W Philips TLD 15 W/08 fluorescent black light blue type, or ten vis lamps (PHILIPS FSLYZ15RR26) being placed horizontally in pairs on each side of the irradiation box at a distance of approximately  $\sim 20\text{ cm}$  from the photocatalytic material. The total light intensity measured on the samples' surface by the ten TLD lamps was on average  $2.6\text{ W m}^{-2}$ , of which 55% ( $1.43\text{ W m}^{-2}$ ) was due to UV. The corresponding total light intensity value for the 10 FSL lamps (vis) was  $7.6\text{ W m}^{-2}$ , of which only the 0.6% was attributed to UV. The irradiation time was set in 45 min. The environmental box was also further equipped with two fans for keeping the temperature stable.

The initial concentration of the air pollutant as well as the adopted temperature and humidity levels into the inner environment of (inside) the glass cell were chosen with the ultimate criterion of approaching the corresponding levels of the same variables in the outdoor real environment. For that purpose the NO concentration was established at approximately  $250\text{ }\mu\text{g m}^{-3}$ , while the temperature and humidity levels were controlled at  $24\text{--}27^\circ\text{C}$  and 35–45% respectively. The  $\text{NO}_x$  concentration was measured with a chemiluminescence  $\text{NO}_x$  analyzer (Environment S.A AC42M) which was connected in line to the outlet of the reactor.

### 2.3. Characterization of manganese-doped $\text{TiO}_2$ nanoparticles

The crystal structure, particle size, morphology, and porosity were examined with an XRD-powder system, SEM, TEM and

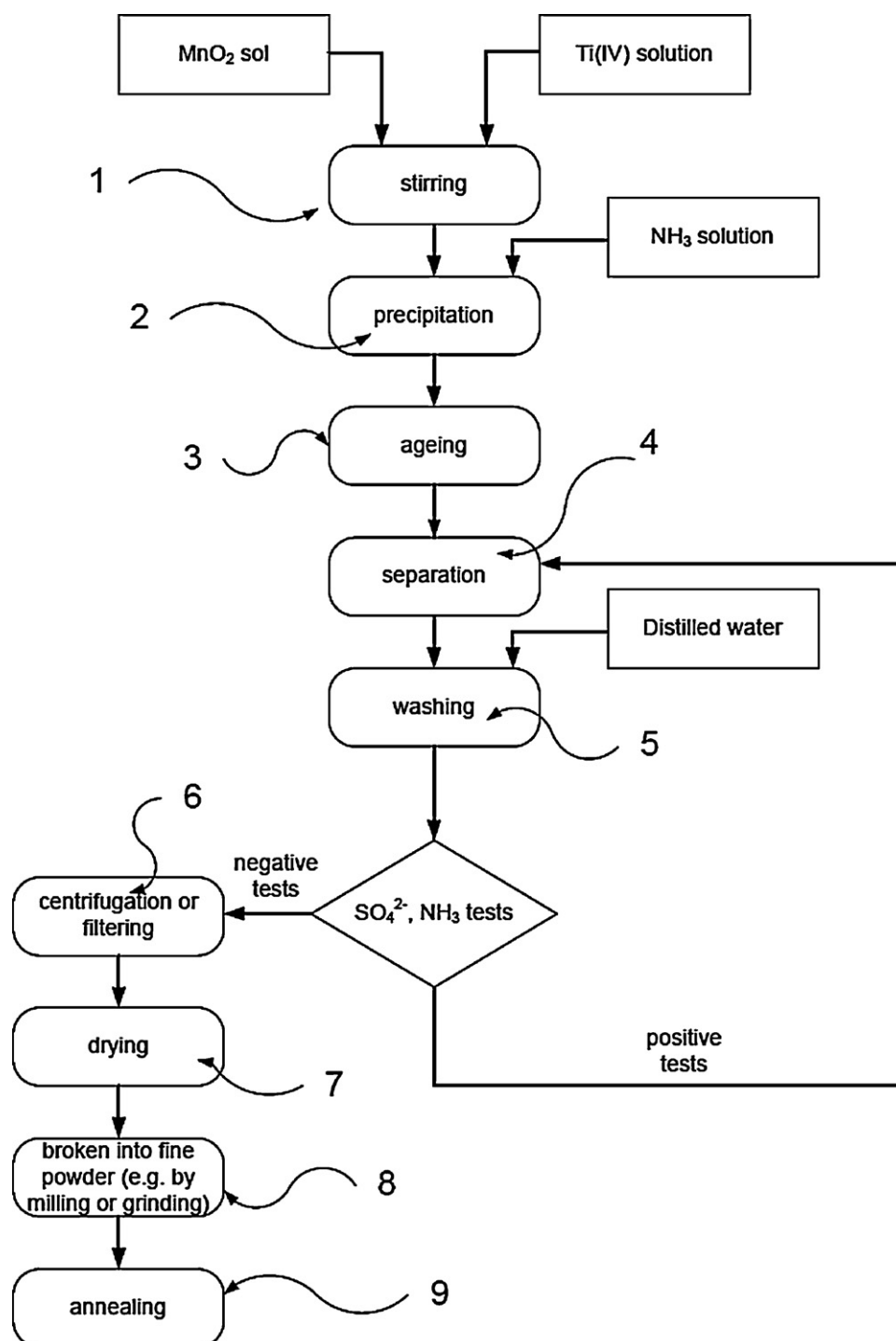


Fig. 1. Synthetic procedure of Mn:TiO<sub>2</sub>.

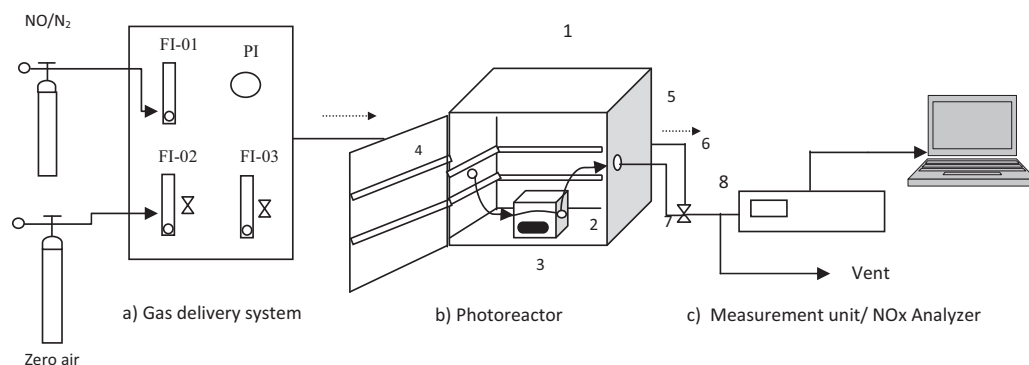
porosimetry respectfully. Powder X-ray diffraction patterns were obtained on a Rigaku D/MAX-2000H rotating anode diffractometer (CuK $\alpha$  radiation) equipped with the secondary pyrolytic graphite monochromator operated at 40 kV and 80 mA over the  $2\theta$  collection range of 20–80°. The scan rate was 0.02° s<sup>-1</sup>. N<sub>2</sub> sorption isotherms were recorded on a commercial volumetric system (Autosorb 1-MP, Quantachrome). Typically, 100 mg of sample was used to record adsorption isotherms. Surface morphology and elemental analysis of the samples were carried out using scanning electron microscopy (SEM) and an energy dispersive spectrometer (EDS) on a JSM-6390LV instrument. The microscopic nanostructures were

observed by TEM on a JEM-2100 instrument equipped with LaB<sub>6</sub> filament, operating at 200 kV.

### 3. Result and discussion

#### 3.1. Structural and optical properties of nanoparticles manganese-doped TiO<sub>2</sub> nanoparticles

The XRD patterns of composite nanostructured powders with different concentrations (0, 0.1, 0.5 and 1 wt%) of Mn:TiO<sub>2</sub> calcined at 700 °C are shown in Fig. 3. For all samples, the anatase phase was



**Fig. 2.** Schematic presentation of the photocatalytic apparatus where FI: flow controller/indicator; PI: pressure indicator; 1: irradiation box; 2: reaction chamber (0.125 m<sup>3</sup>); 3: Mn:TiO<sub>2</sub> sample (0.05525 m<sup>2</sup>); 4: UV-vis or UV Lamps (10); 5: fans (2); 6: by-pass line; 7: base line; 8: sampling line.

detected with corresponding values at  $2\theta = 25.3^\circ$  for plane (101),  $38^\circ$  for (004), and  $48^\circ$  for (200). The grain size ( $D$ ) for all samples was in the range of 20–30 nm and it was determined from the full width at half maximum (FWHM) of the (101) anatase peak according to the Scherrer's formula,  $D = k\lambda / w \cos \theta$  [where  $k$  is the shape factor ( $\sim 0.9$ ),  $\lambda$  the X-ray wavelength (0.15418 nm), and  $\theta$  the diffraction angle].

Fig. 4 shows the effect of calcination temperature on the structure of the XRD patterns of 0.1% Mn:TiO<sub>2</sub>. A monotonic increase of the intensity of the peaks in the temperature range of 200–700 °C was obtained. All samples in the above range of calcination temperatures were monophasic with only the anatase polymorph TiO<sub>2</sub> been detected.

The BET (Brunauer–Emmet–Teller) surface area of the Mn:TiO<sub>2</sub> composites with different concentration of Mn (0.1%, 0.5% and 1%) calcined at 700 °C for 3 h was calculated at 24, 30 and 38 m<sup>2</sup> g<sup>−1</sup>, respectively.

Absorption studies on the said material as a function of Mn concentration presented in a recent publication of Papadimitriou et al. [35] elaborate on the role of Mn in the induced catalytic behavior of 0.1% Mn:TiO<sub>2</sub>. It was found that at an optimum dopant concentration of 0.1% the material absorption is maximum, an observation that is in line with past reports by Ullah and Dutta [36] on a Mn doped ZnO system. It is believed that at very low concentrations (<0.5%) Mn, taking interstitial or substitutional sites in TiO<sub>2</sub> matrix,

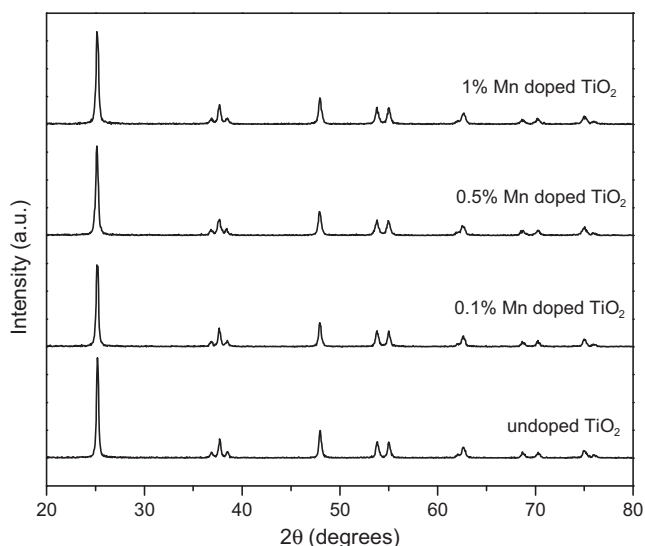
is contributing to the enhancement of the material absorption with a corresponding red-shift. At higher concentrations significant structural changes of TiO<sub>2</sub> may take place with manganese, in the form of Mn<sup>2+</sup>, reacting preferably with oxygen to form MnO<sub>x</sub>. Thus, we observed a reduction in the optical absorption with further increase in the dopant concentration and the existence of an optimum at 0.1% Mn doping.

The textural features of the samples were investigated with scanning electron microscopy. Representative images of composites in different concentrations for 0%, 0.1%, 0.5% and 1% Mn:TiO<sub>2</sub> are shown on Fig. 5 demonstrating no specific morphology changes apart of a trend for surface particle agglomeration. Corresponding energy dispersive X-ray spectroscopy (EDX) indicated the presence of Mn and TiO<sub>2</sub> nanocomposites. TEM images of the 1% Mn:TiO<sub>2</sub> composite confirmed the agglomeration of particles with size estimated about 20–30 nm (Fig. 6).

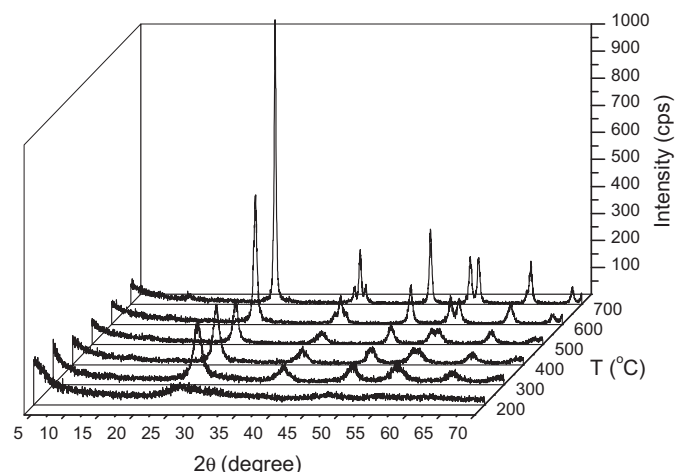
### 3.2. Photocatalytic properties

#### 3.2.1. Degradation of Methylene Blue (MB)

The photocatalytic activities of undoped and Mn:TiO<sub>2</sub> were carried out using UV light and visible light. MB was used as the test contaminant, a dye extensively used as an indicator for the study of photocatalytic activities [37,38] exhibiting absorption peak (s) in the visible range. Thus, degradation can be easily monitored by optical absorption spectroscopy analyzing the evolution of the 675 nm MB peak. The recorded decolorization of MB under UV and vis light in the absence of photocatalytic material (shown in Figs. 7 and 8) might be due to the interaction of MB with OH•



**Fig. 3.** Powder XRD patterns of undoped and Mn doped TiO<sub>2</sub> samples in different concentration of manganese.



**Fig. 4.** Powder XRD patterns in different calcinations temperatures of 0.1% Mn:TiO<sub>2</sub>.



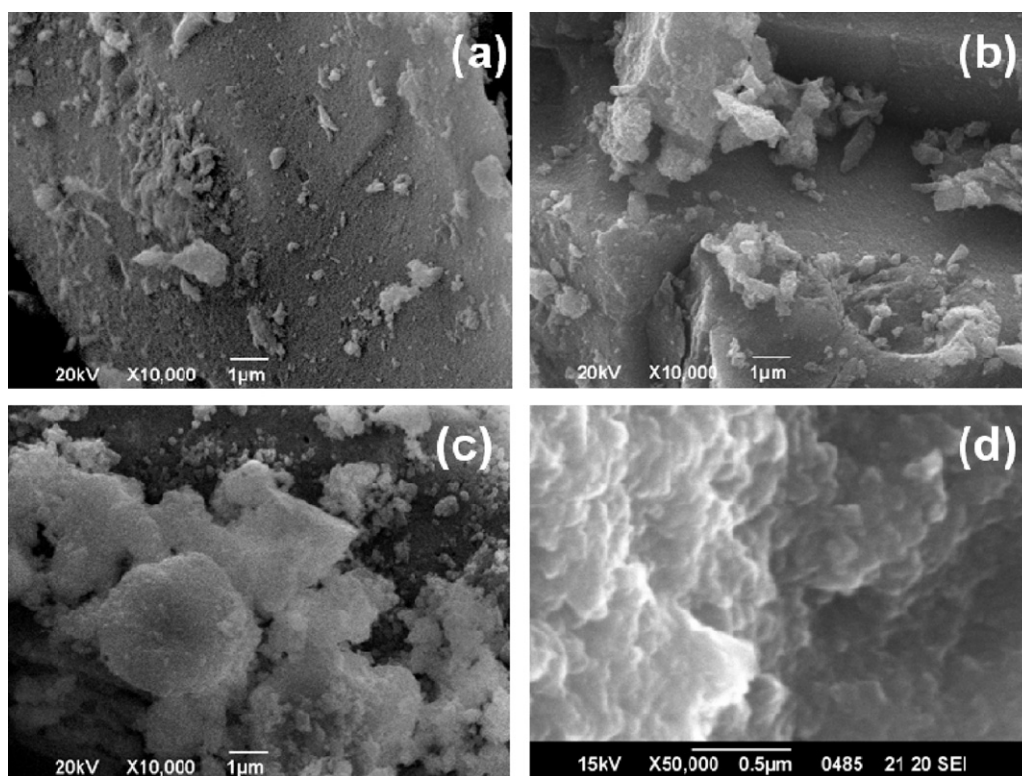


Fig. 5. Scanning electron microscopy (SEM) images of Mn:TiO<sub>2</sub> powders in different concentrations (a: undoped TiO<sub>2</sub>, b: 0.1%, c: 0.5%, d: 1%).

radicals originated from water [39]. In our tests we first studied the photodegradation of MB in pure TiO<sub>2</sub> and subsequently in Mn:TiO<sub>2</sub> in different loadings (0.1%, 0.5%, 1%) by irradiating a mixture of photocatalyst and MB with UV light. It was observed that doped TiO<sub>2</sub> decolorizes MB in a monotonic way with a decolorization decreasing as increasing the loading of Mn doping. Superior decomposition efficiency for the 0.1% Mn material with respect of all other compositions was recorded. However in all tests under UV light we did not record any significant improvement on decolorization efficiencies between the 0.1% Mn doped and the undoped TiO<sub>2</sub> materials (Fig. 7). In contrast with the above, the photo-degradation rate of MB with 0.1% Mn:TiO<sub>2</sub> irradiated with visible light was found to be faster than with other loadings as well as that of the undoped TiO<sub>2</sub> as shown in Fig. 8. It was recorded that it only takes 20 min for 0.1% Mn:TiO<sub>2</sub> to decolorize as much as 70% of MB while in the presence

of undoped TiO<sub>2</sub> the decolorization was only 40% for the same time scale (Figs. 9 and 10).

### 3.2.2. Degradation of NO<sub>x</sub>

The photocatalytic removal of NO from 0.1% of Mn:TiO<sub>2</sub> powder in calcareous matrices containing 5% and 10% of the photocatalytic material was evaluated under both UV and vis irradiation. Photocatalytic experiments were conducted in order to distinguish the pure UV or vis photocatalytic effect of the TiO<sub>2</sub> powders from the NO<sub>x</sub> reduction due to its adsorption on the cell walls and/or its radiation photolysis. Thus, initially, blank tests were carried out by polluting the reactor with NO in the absence of the photocatalyst without

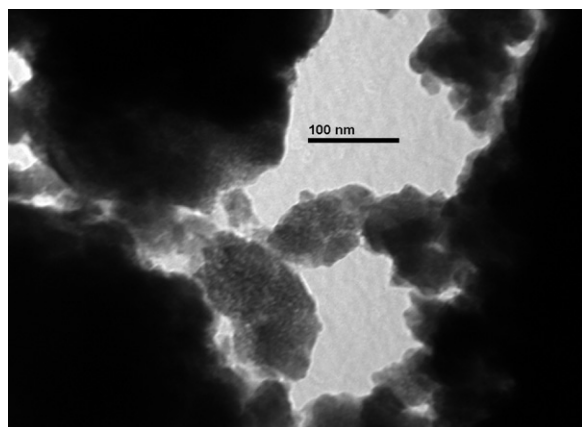


Fig. 6. TEM image of the Mn:TiO<sub>2</sub> (1:100).

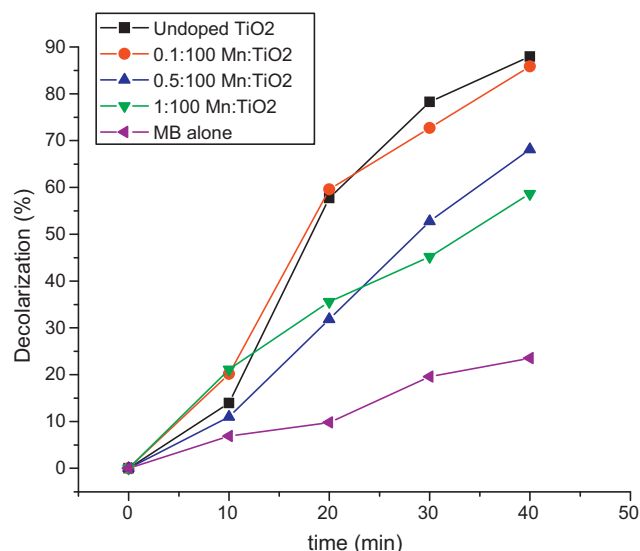


Fig. 7. Degradation of Methylene Blue (MB) under UV light.

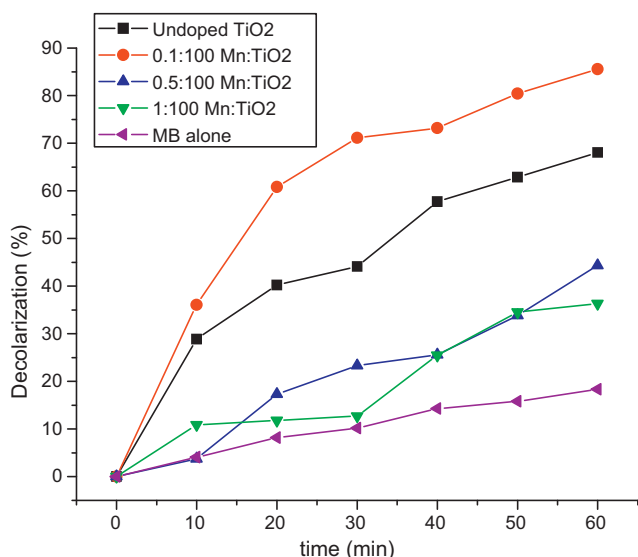


Fig. 8. Degradation of Methylene Blue under visible light.

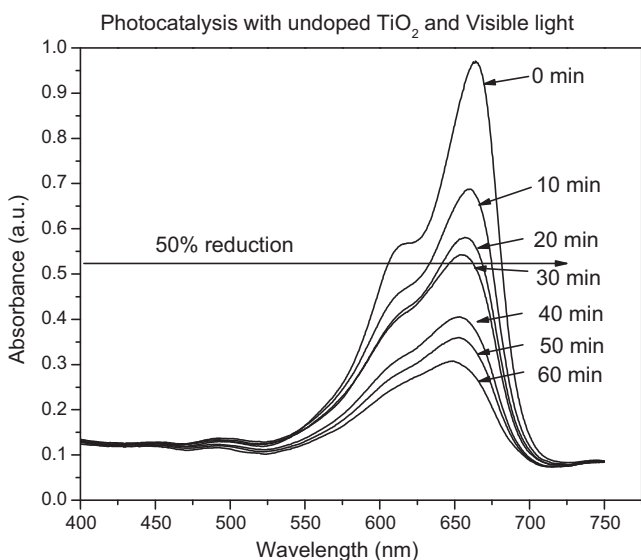


Fig. 9. Photo-reduction of MB with undoped TiO<sub>2</sub> and visible light.

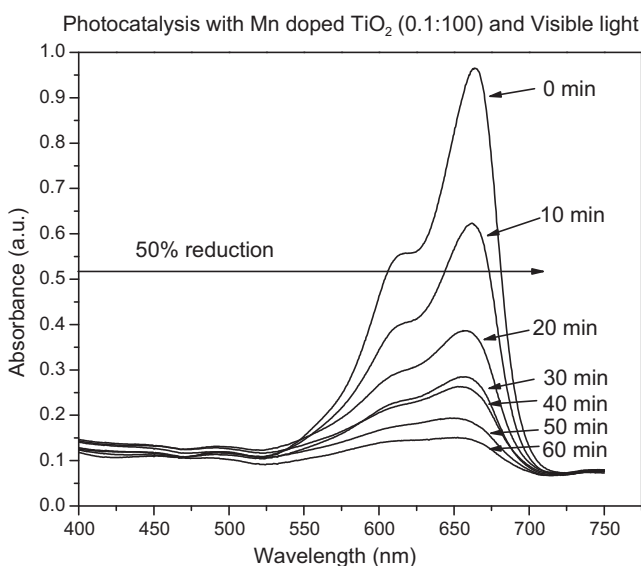


Fig. 10. Photo-reduction of MB with Mn:TiO<sub>2</sub> and visible light.

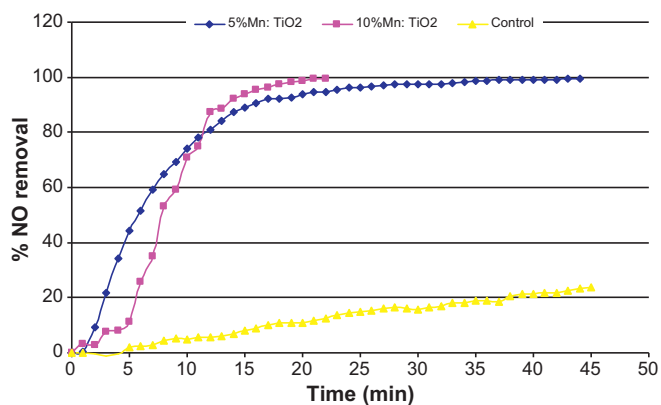


Fig. 11. %NO degradation under UV illumination.

and with irradiation respectively. Then the same experiments were carried out in the presence of the photocatalysts.

Photodegradation graphs of the different samples under UV and vis irradiation are shown in Fig. 11 and Fig. 12 respectively. As it can be seen from both figures, immediately after the irradiation of the sample (Time 0), a significant decrease in the NO concentration is observed which reveals the immediate response of the photocatalytic system and gives the size of the photocatalytic performance of the materials. A slightly better photocatalytic ability was monitored by the 10% Mn:TiO<sub>2</sub> powder under both UV and vis with respect to the 5%. Additionally, it is shown (Fig. 11) that the 10% sample removes more than 99% of NO from the system after 25 min while the 5% needs more than 40 min to remove the same amount. Independent studies on the same materials and concentrations carried out by the JRC at ISPRA [40] have shown identical photocatalytic activities, thus, underlining the effectiveness of these Mn:TiO<sub>2</sub> powders for in-door depollution application utilizing artificial lighting.

The reduction of NO during blank tests (substrate and filler without (w/o) TiO<sub>2</sub>) is attributed to competitive depletion mechanisms such as adsorption of NO onto the chamber's wall area and to photolysis. Calculations have shown that both these mechanisms did not contribute to more than 32% of the total NO removal during the photocatalytic experiments (Tables 1 and 2).

The above results have been further analyzed in order to quantify the photocatalytic performance of the TiO<sub>2</sub> materials under UV or vis irradiation. The performance of the materials was evaluated via the calculation of the following parameters: (a) the percentage

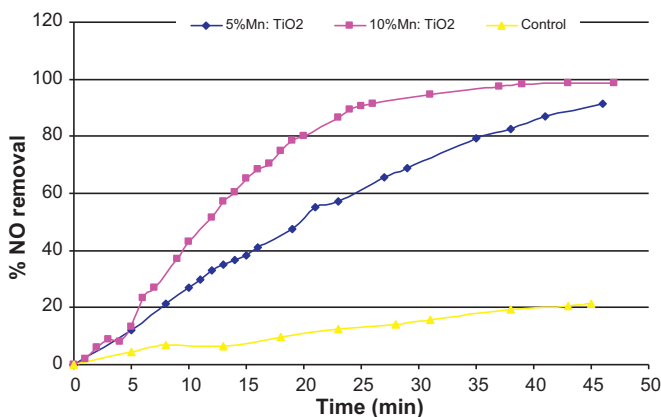


Fig. 12. %NO degradation under vis illumination.

**Table 1**

Material's photocatalytic activity results under UV irradiation.

Material	% NO removal	% NO photocatalytic destruction (PPD)	Photocatalytic destruction rate (DR) NO ( $\mu\text{g m}^{-2} \text{s}^{-1}$ )
Substrate	32 (Removal due to UV photolysis and sink effects on reactor's wall)	–	–
Calcareous filler w/o Mn:TiO <sub>2</sub>	35	–	–
Calcareous filler with 5% Mn:TiO <sub>2</sub>	99.5	67.5	0.151
Calcareous filler with 10% Mn:TiO <sub>2</sub>	99.7	67.7	0.185

**Table 2**

Material's photocatalytic activity results under vis irradiation.

Material	% NO removal	% NO photocatalytic destruction (PPD)	Photocatalytic destruction rate (DR) NO ( $\mu\text{g m}^{-2} \text{s}^{-1}$ )
Substrate	30 (Removal due to vis photolysis and sink effects on reactor's wall)	–	–
Calcareous filler w/o Mn:TiO <sub>2</sub>	30.9	–	–
Calcareous filler with 5% Mn:TiO <sub>2</sub>	92.4	62.4	0.129
Calcareous filler with 10% Mn:TiO <sub>2</sub>	98.8	68.8	0.130

photocatalytic degradation (PPD) and (b) the destruction rate (DR) according to the following Eqs. (1)–(2).

$$\text{PPD}(\%) = \left[ \frac{C_{\text{in}} - C_{\text{fin}}}{C_{\text{in}}} \right] \times 100 \quad (1)$$

$$\text{DR}(\mu\text{g}/\text{m}^2\text{s}) = (C_{\text{in}} - C_{\text{fin}}) \times \frac{V}{A \times t} \quad (2)$$

where,  $C_{\text{in}}$ ,  $C_{\text{fin}}$  the initial and final NO concentrations,  $A$  the sample area and  $V$  the volume of the reactor.

The destruction rate (DR) of NO (Tables 1 and 2) provides a more accurate measure of the photocatalytic activity of the material in comparison with the photocatalytic decomposition parameter PPD (%) owing to the fact that the DR is taking into consideration the initial concentration of the air pollutant as well as the sample's area and the irradiation time and is expressed as  $\mu\text{g}$  of converted NO per  $\text{m}^2$  of material per second of irradiation. As far as the 5% and 10% Mn:TiO<sub>2</sub> materials under UV irradiation are concerned the DR have been calculated to be 0.151 and 0.185 respectively. The corresponding experiments under vis irradiation presented slightly lower values calculated to be 0.129 and 0.130 respectively.

Currently, research is focusing on the improvement of the above synthetic procedure which, combined with further characterization of the composites nanostructure and the use of other dopand (transition metal or non metal) materials, form an extensive study on the degradation mechanism of additional volatile organic compounds (formaldehyde, toluene, benzene) and inorganic compounds (NO<sub>x</sub> and CO) under UV and visible light for out-door and in-door applications.

#### 4. Conclusions

Nanocrystalline Mn:TiO<sub>2</sub> powder of crystallite size of 20–30 nm can be synthesized by a straight forward and inexpensive procedure. The starting chemicals applied for this process are easily accessible, and are cheaper than alternative organic precursors leading to the same metal oxides powders. Anatase was found to be the major phase in all samples. FT-IR/vis adsorption has shown enhanced sub-band-gap absorption as a function of Mn concentration. The photocatalytic activity on the MB degradation of 0.1% Mn:TiO<sub>2</sub> was higher than the other photocatalytic material with different concentrations (0.5% and 1%) both under UV and visible light. Corresponding tests on the NO removal by calcareous bases containing 5% and 10% of the 0.1% Mn:TiO<sub>2</sub> powder material have

shown a 99% reduction rate under UV and vis light exposure in 25 and 45 min respectfully. Corresponding maximum reduction rates were 0.185 and 0.130 (for the 10% content in the calcareous base) of the 0.1% Mn:TiO<sub>2</sub> powder.

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